



**Baffin Bay Volunteer Water Quality Monitoring Study: Synthesis of May
2013-July 2015 Data**

Annual Report

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Executive Summary

The goal of this study is to quantify spatial-temporal distributions of key water quality parameters in Baffin Bay, relying on volunteer citizen scientists for data collection. Sample collection began in May 2013 from 9 sites throughout Baffin Bay and continues to present. At the beginning of the study period, Baffin Bay was experiencing a significant, prolonged drought. In early fall 2013, the drought began to lessen and precipitation patterns more in accordance with long-term monthly averages developed. By spring 2015, El Niño-like conditions developed, which led to several periods of intense rainfall in the Baffin Bay watershed. Salinities were very high at the start of the sampling period (>60-70), but decreased over the course of the study, with the decrease accelerating in spring 2015 concurrent with the heavy rainfall. Results show distinct seasonal and spatial patterns in several important water quality parameters. Of particular interest is chlorophyll *a*, an indicator of algal biomass. Chlorophyll tended to peak in spring and reached its lowest levels in late fall-winter. Interestingly, chlorophyll concentrations were nearly 2-fold higher in winter 2014-2015 compared to winter 2013-2014, concurrent with water temperatures that were ca. 5°C warmer in winter 2014-2015. Findings show persistent, excessive levels of chlorophyll in a large portion of Baffin Bay for much of the study period, often exceeding TCEQ criteria for estuaries and thus indicating a nutrient enriched system. Only recently (spring 2015) has there been an obvious system-wide decrease in chlorophyll levels, likely due to the flushing effects associated with the recent rains. Inorganic nutrient concentrations (ammonium, nitrate, orthophosphate) in surface waters were high at the beginning of the record, but decreased thereafter and remained at very low levels until spring 2015, when concentrations increased as a result of runoff. Despite typically low surface water concentrations, ammonium and orthophosphate concentrations were often elevated in near bottom waters during summer, indicating continued availability of nutrients via recycling to support algal growth. Dissolved organic nitrogen levels were very high in the system, with system-wide average concentrations consistently exceeding 48 µM throughout the study. Dissolved organic carbon levels were also very high in spring-early summer of 2013 (718 to >2400 µM) and declined thereafter, but still consistently exceeded 400 µM. Dissolved oxygen levels showed a distinct seasonal cycle that is temperature dependent. Undersaturated dissolved oxygen levels were noted each summer in Baffin Bay, indicative of high biological oxygen demand that is supported by labile organic matter. These results highlight the influence of

seasonal factors in terms of determining water quality conditions, as well as the influence of drought-wet cycles. Additional data collections through 2017 will be crucial for determining the longer-term influence of wet periods, such as those associated with the recent initiation of El Niño conditions.

Acknowledgements

First and foremost, we thank the many volunteers who dedicated time and resources to the collection of this water quality data in Baffin Bay. This study would not have been possible without their efforts. We are also grateful to the entities who currently fund the study, including Celanese Corporation and the NOAA/Texas GLO Coastal Management Program, as well as past funding entities including Kleberg County, the Coastal Bend Bays & Estuaries Program, Coastal Conservation Association and the Saltwater Fisheries Enhancement Association. Finally, we thank our partners at the Coastal Bend Bays & Estuaries Program for their ongoing support and interest.

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Introduction

Cultural eutrophication is a major environmental threat facing coastal ecosystems worldwide (Nixon 1995; Diaz and Rosenberg 2008). Over the past 50 years, there has been a substantial increase in nutrient loading to the coastal zone, resulting in growing expression of symptoms such as persistent algal blooms, hypoxia/anoxia formation, and microbial pathogen growth among others (Nixon 1995; Boesch 2002; Rabalais et al. 2009). These symptoms often have deleterious consequences for ecosystem structure and function, resulting in such visible effects as fish kills and other animal mortalities, alteration of food webs and economic losses (Diaz and Rosenberg 1995; Boesch 2002). The most recent synthesis of data from the U.S. indicates that as of 2007, at least 30% of estuaries were considered moderately to highly eutrophic, with eutrophication pressures expected to grow in 65% of estuaries over the next decade (Bricker et al. 2007). Unfortunately Texas estuaries have been poorly represented in national eutrophication assessments including the aforementioned report, largely due to lack of sampling efforts and data coverage. Nonetheless, there is growing concern fueled by public observations and recent scientific assessments that several systems in South Texas are indeed undergoing significant eutrophication. One example is Baffin Bay, which represents critical habitat for several economically- and ecologically-important fish species and is popular with recreational fishermen.

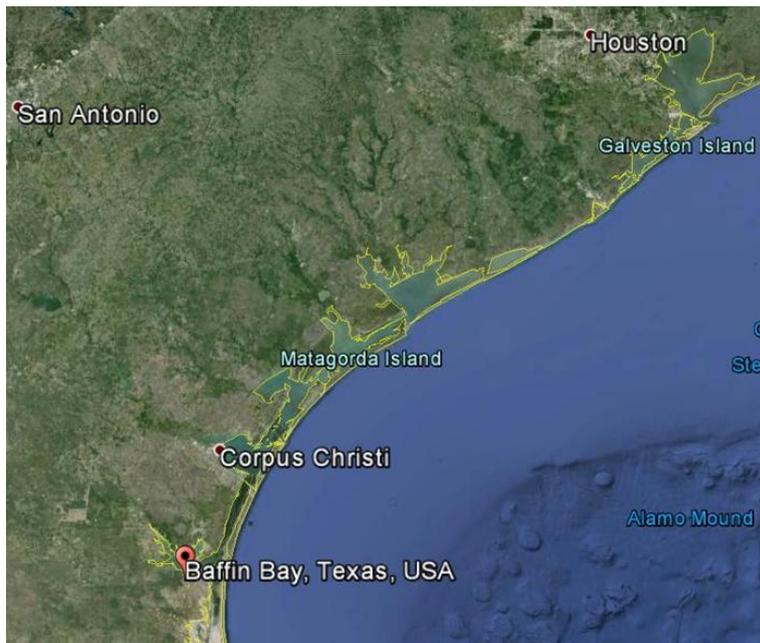
In the past 1-2 decades, growing expression of symptoms of eutrophication such as hypoxia and dense algal (phytoplankton) blooms have been noted in Baffin Bay. Hypoxia and excessive phytoplankton growth, which are quite possibly intricately linked, are concerning because of their potential effects on ecosystem health and fisheries in coastal embayments. For instance, hypoxia has been linked to several large fish kills in Baffin Bay over the past five years (unpubl. Texas Parks & Wildlife Spills & Kills Team reports). Hypoxia formation tends to occur during warm summer-fall months, often following freshwater pulses that inject allochthonous nutrients and organic matter and induce stratification in the bays (unpubl. Texas Parks & Wildlife Spills & Kills Team reports). Co-occurrence of phytoplankton blooms and hypoxia have been noted in Baffin Bay as well (unpubl. Texas Parks & Wildlife Spills & Kills Team reports), and overall phytoplankton biomass frequently exceeds state criteria, raising concerns about the potential role of nutrient-laden runoff (Montagna and Palmer 2012; this study). For instance, Baffin Bay has experienced prolonged, dense blooms of the brown tide organism, *Aureoumbra lagunensis*, since

1990 (Buskey et al. 1997; Buskey et al. 2001). More recently, a fish kill occurred in 2010 and coincided not only with hypoxia, but also with a dense phytoplankton bloom of the dinoflagellate *Pyrodinium bahamense* and the diatom *Thalassiothrix sp.* (unpubl. Texas Parks & Wildlife Spills & Kills Team report).

Using data obtained primarily from TCEQ quarterly sampling, Montagna and Palmer (2012) documented a long-term increase in Kjeldahl nitrogen, nitrate and orthophosphate in Baffin Bay. Ammonium, chlorophyll *a* and nitrate also regularly exceeded state criteria for water quality standards in a number of years. While state agency sampling efforts in Baffin Bay have been valuable for documenting long-term water quality changes in the system, their limited spatial-temporal coverage hinders determination of the timing and location of symptoms of water quality degradation, and also preclude determination of the main cause(s) of water quality degradation in the system. Here results are presented from the first two years of a volunteer water quality monitoring study, the goals of which are to quantify spatial-temporal distributions of key water quality parameters in Baffin Bay, and to increase our understanding of the drivers of water quality change in this system.

Methods

Study location – Baffin Bay is a shallow ($\leq 2\text{-}3$ m depth) South Texas coastal embayment



adjacent to the Laguna Madre (Figure 1). Residence time of water in Baffin Bay exceeds 1 year due to minimal tidal influence and freshwater inflows, and the system is prone to hypersaline conditions due to evaporation exceeding precipitation (Shormann 1992). Circulation in Baffin Bay is primarily driven by winds.

Figure 1. Map of Baffin Bay, located ~50 km southwest of Corpus Christi, TX.

Meteorological data – Monthly mean precipitation data from the Naval Air Station Kingsville was obtained from the National Climatic Data Center. Using data from January 1973 through June 2014, monthly mean precipitation was calculated. The monthly deviation from this long-term monthly mean was then calculated, and is heretofore referred to as precipitation “anomaly”.

Sample collection – Water samples were collected on a monthly basis from May 2013 through July 2015 at 5-12 sites in Baffin Bay, weather permitting (Table 1). This included 9 permanent sites (Figure 2) and 3 additional sites in the head of each tributary (Cayo del Grullo, Laguna Salada, Alazan) following significant rain events. Water samples were collected by volunteer citizen scientists. In order to qualify for this program, volunteers had to undergo rigorous training in the lab of Dr. Michael Wetz (Texas A&M University-Corpus Christi) and demonstrate competency in field sample collection (documentation retained in Wetz lab). At

Sampling Date	Sites
May 15 th and 16 th , 2013	8
June 13 th , 2013	9
July 18 th , 2013	12
August 15 th , 2013	9
September 18 th , 2013	9
October 18 th , 2013	5; adverse weather at other sites
November 15 th , 2013	9
December 13 th , 2013	9
January 16 th , 2014	9
February 13 th , 2014	9
March 19 th , 2014	9
April 16 th , 2014	9
May 16 th , 2014	12
June 27 th , 2014	9
July 17 th , 2014	9
August 14 th , 2014	9
September 26 th , 2014	9
October 16 th , 2014	9
November 20 th , 2014	9
December 18 th , 2014	9
January 29 th , 2015	9
February 19 th , 2015	9
March 18 th , 2015	9
April 16 th , 2015	9
May 20 th , 2015	9
June 24 th , 2015	9
July 23 rd , 2015	9

each site, a profile of salinity, temperature, conductivity, dissolved oxygen and pH was obtained by lowering a YSI ProPlus sonde at 0.5 m increments through the water column. Surface and near bottom discrete water samples were collected in a Van Dorn sampling device and transferred to acid-washed amber polycarbonate bottles. Bottles were stored on ice until return to a shore-based facility where processing of samples occurred.

Table 1. Sampling events completed to date.

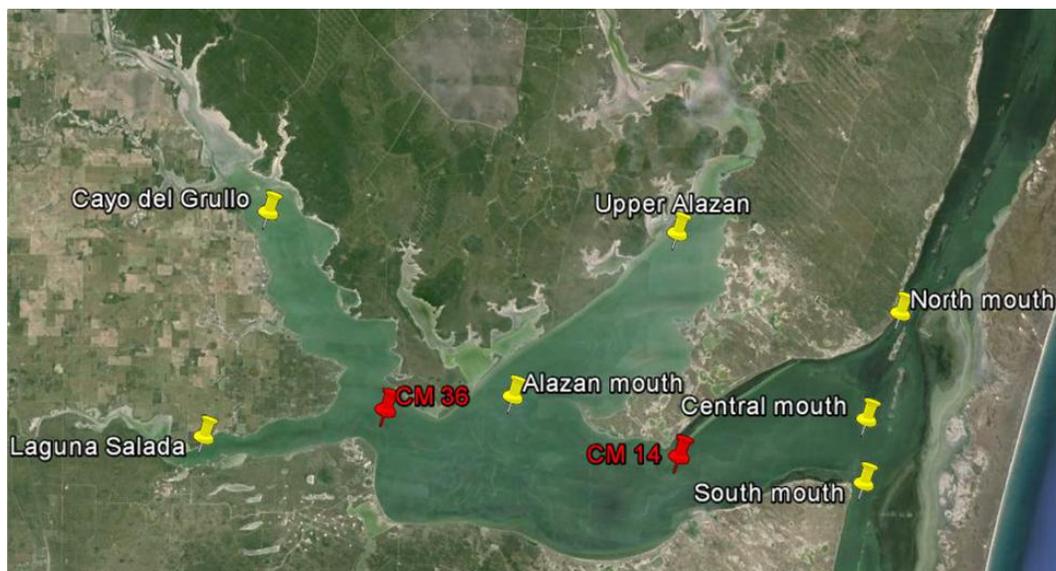


Figure 2. Map of permanent sampling locations in Baffin Bay. Red markers indicate two sites that are visited as part of TCEQ’s quarterly monitoring program.

Sample analyses – Chlorophyll *a* was determined from samples collected on, and extracted from Whatman GF/F filters (nominal pore size 0.7 μm). Chlorophyll was extracted using 90% acetone and analyzed fluorometrically. Inorganic nutrients (nitrate + nitrite (N+N), nitrite, silicate, orthophosphate, ammonium) were determined in the filtrate of water that passed through GF/F filters using a Seal QuAAtro autoanalyzer. Dissolved organic carbon (DOC) and total dissolved nitrogen (TDN) were determined in the filtrate of water that passed through GF/F filters using a Shimadzu TOC-V analyzer with nitrogen module. Dissolved organic nitrogen (DON) was estimated as the difference between TDN and inorganic nitrogen. Complete methodological details on wet chemical and YSI analyses can be obtained from Dr. Wetz.

Results

At the beginning of the study period, Baffin Bay was experiencing a significant, prolonged drought (Fig. 3). In early fall 2013, the drought began to lessen and precipitation patterns more in accordance with long-term monthly averages developed (Fig. 3). By spring 2015, El Niño-like conditions developed, which led to several periods of intense rainfall in the Baffin Bay watershed.

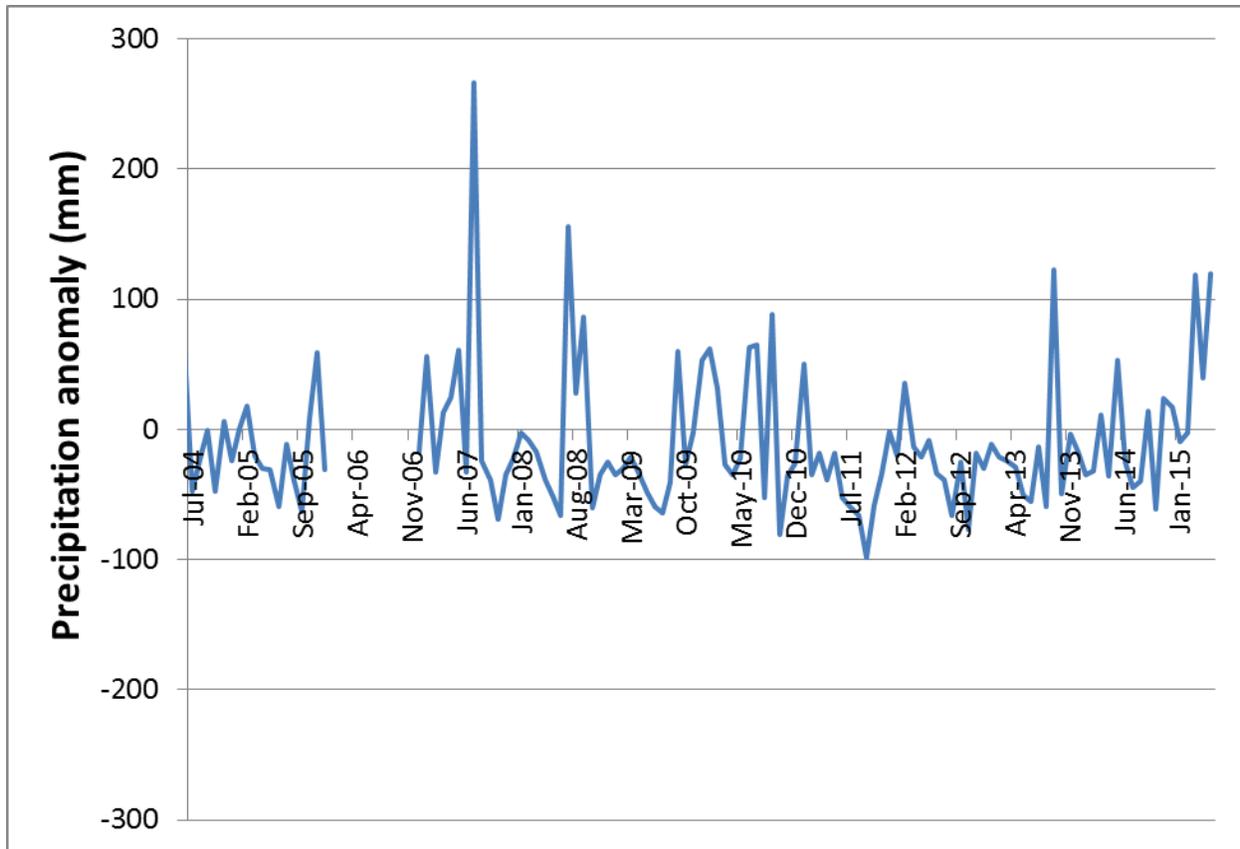


Figure 3. Precipitation anomaly (deviation from 1974-2014 monthly mean) at Kingsville Naval Air Station from 2004 to 2014.

Physical setting – Water temperature followed a distinct seasonal pattern, with temperatures increasing in late winter-early spring, peaking during summer, and then decreasing in early fall (Fig. 4). It is interesting to note that water temperature was ~5°C higher in winter 2014-2015 compared to winter 2013-2014 (Fig. 4). There were no obvious water temperature differences between sites (data not shown). Salinity was very high at the start of the sampling period in May 2013, exceeding 70 in the upper Alazan and Laguna Salada (Fig. 5). Highest salinity was consistently recorded in western Baffin Bay, particularly in the tributaries, while lowest salinity was observed near the mouth of Baffin Bay. Salinity tended to decrease over the course of the study, with the decrease accelerating in spring 2015 as a result of heavy rainfall in the watershed (Fig. 5). In early summer of both 2013 and 2014, intrusion of lower salinity water from Laguna Madre was apparent, particularly in the eastern portion of Baffin Bay (Fig. 5,6). This influence of lower salinity Laguna Madre water did not extend into western Baffin Bay however. Strong

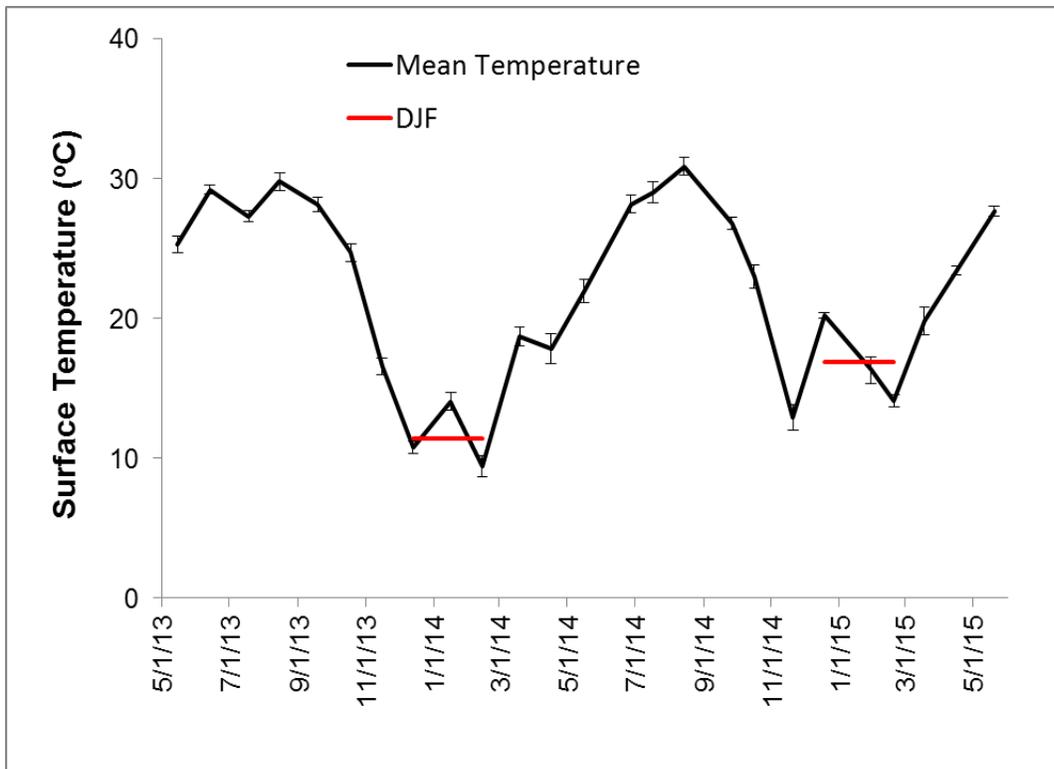


Figure 4. Surface temperature in Baffin Bay.

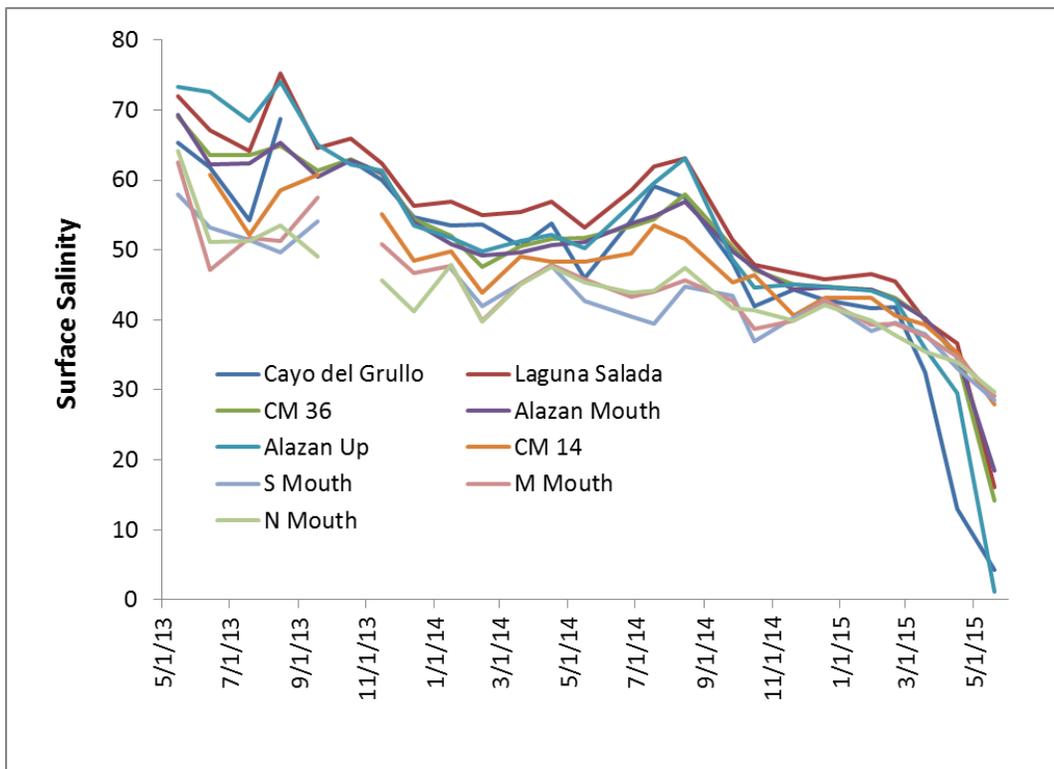


Figure 5. Surface salinity in Baffin Bay.

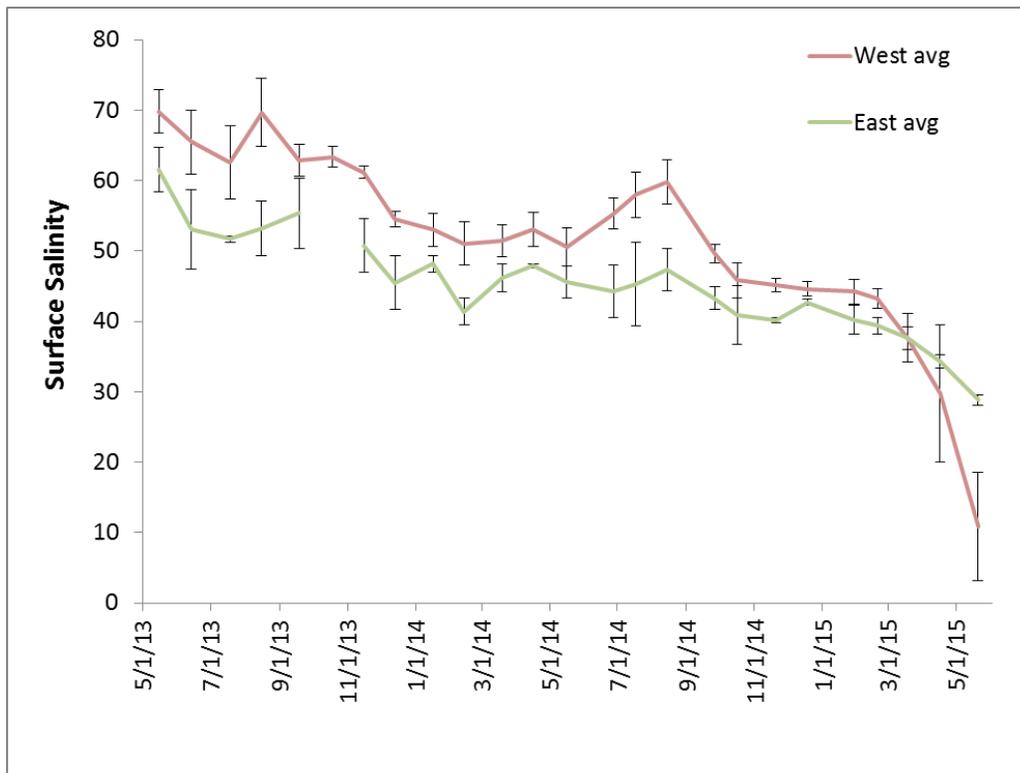


Figure 6. Surface salinity in western compared to eastern Baffin Bay. “West” includes upper Alazan, lower Alazan, CM 36, Laguna Salada and Cayo del Grullo sites. “East” includes CM 14, north/central/south mouth sites.

salinity stratification (i.e., higher salinity in subpycnocline waters than surface waters) of the water column was observed in summer 2013, but was not as noticeable in summer 2014 (Fig. 7).

Biological-chemical dynamics – Chlorophyll *a* concentrations tended to be very high in Baffin Bay during this study, exceeding the TCEQ criteria (11.6 $\mu\text{g/l}$) in 175 of 220 samples (80%). Using a slightly more relaxed National Coastal Condition Report for “poor” condition (20 $\mu\text{g/l}$; NCCR 2012), chlorophyll *a* was still in excess in 113 of 220 sample collections (51%). A distinct seasonal pattern was observed in terms of chlorophyll *a*, with highest concentrations found in spring-summer, decreasing through fall-winter (Fig. 8). Interestingly, chlorophyll *a* concentrations were nearly 2-fold higher in winter 2014-2015 compared to winter 2013-2014, coincident with water temperatures that were ca. 5°C warmer in winter 2014-2015 (Figs. 4, 8). During the observed intrusion of lower salinity Laguna Madre water in early summer of both

years, chlorophyll tended to decrease in the eastern portion of Baffin Bay, but remained elevated in the western portion (Fig. 9).

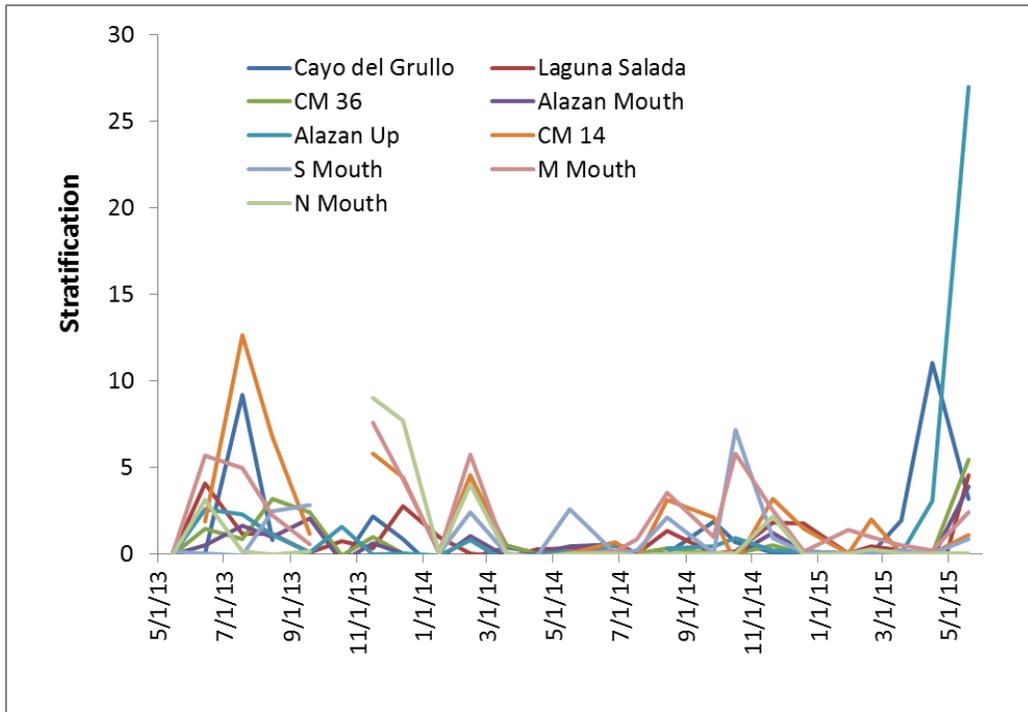


Figure 7. Stratification index (=Bottom minus surface salinity) in Baffin Bay.

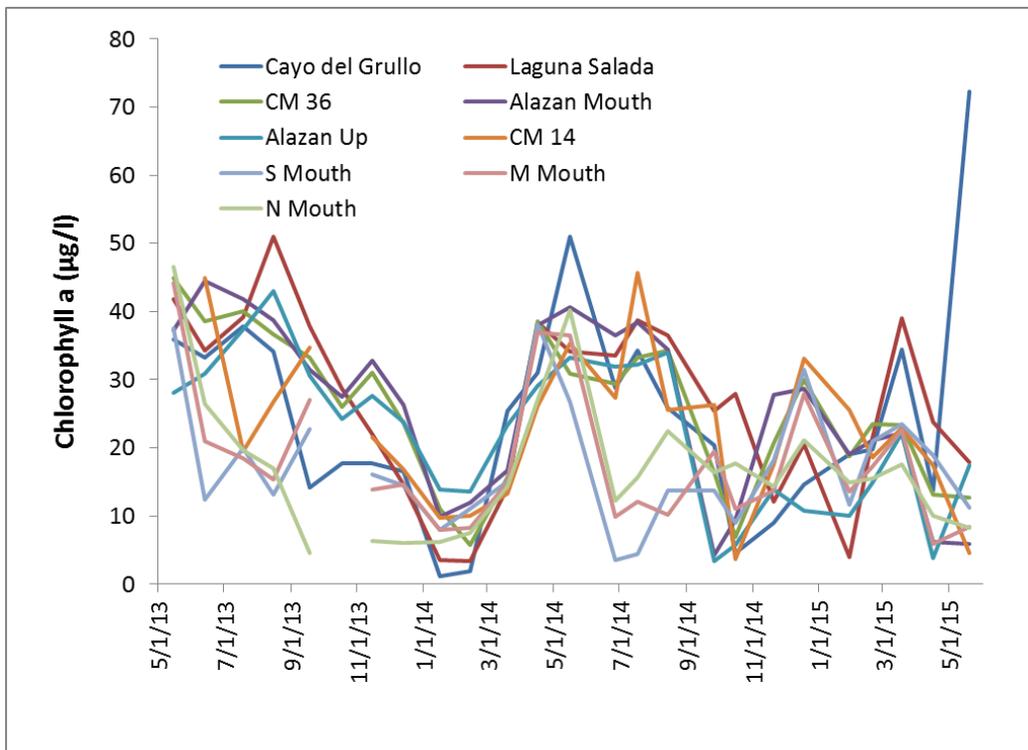


Figure 8. Chlorophyll a in Baffin Bay.

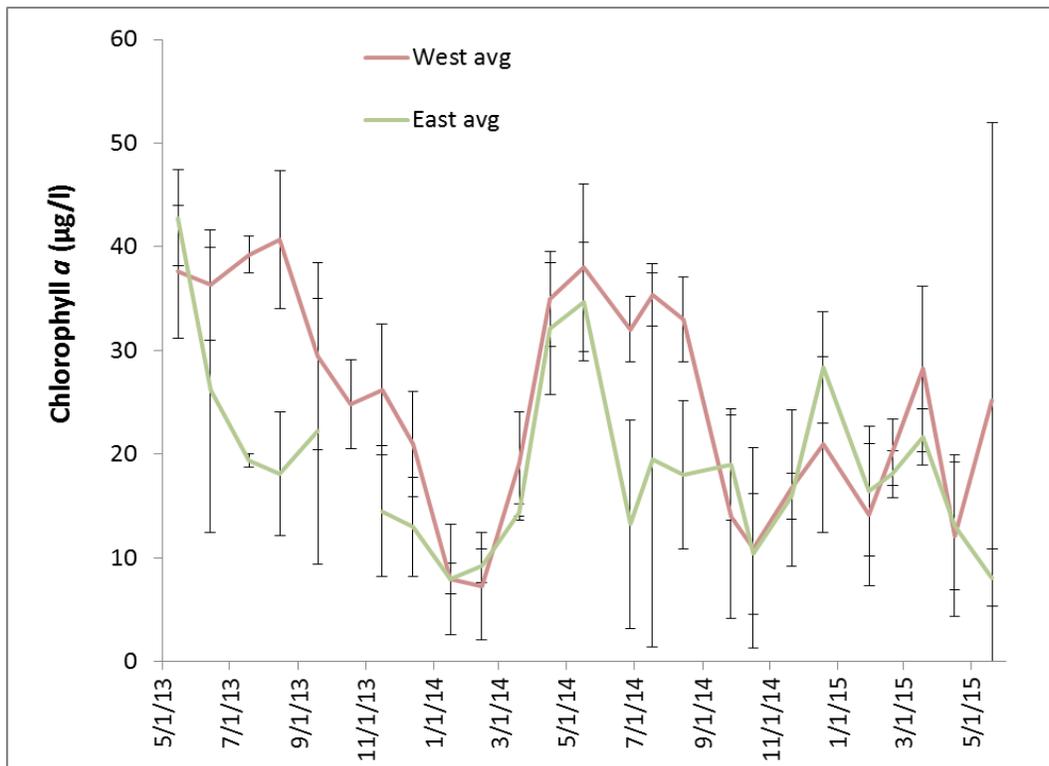


Figure 9. Chlorophyll *a* in western compared to eastern Baffin Bay. “West” includes upper Alazan, lower Alazan, CM 36, Laguna Salada and Cayo del Grullo sites. “East” includes CM 14, north/central/south mouth sites.

In terms of the inorganic nutrients that may support phytoplankton growth, surface nitrate plus nitrite (N+N) concentrations exceeded 5 µM at the beginning of the study period at all sites except Cayo del Grullo (4.4 µM), and the north and central sites at the mouth (< 1 µM) (Fig. 10). Very high N+N concentrations (>35 µM) were noted at both the upper Alazan site and south site at the mouth. N+N concentrations decreased after May 2013 and were generally <1-4 µM thereafter through February 2015. One relatively high N+N value (6.9 µM) was observed at the upper Alazan site in March 2015, while N+N at other sites was <0.3 µM (Fig. 10). In April 2015, N+N was 15.9 µM at the Cayo del Grullo, but <0.6 µM at the other sites (Fig. 10). Finally, in May 2015 two relatively high N+N values were observed, 6.8 µM at CM 36 and 28.3 µM at the upper Alazan site (Fig. 10), while other sites had N+N <1.5 µM. N+N concentrations were elevated in near bottom waters in summer 2013, but not summer 2014. Relatively high

N+N bottom water concentrations were observed in several locations in spring 2015 (Fig. 11). Surface ammonium concentrations were high in July-August 2013, with highest concentrations

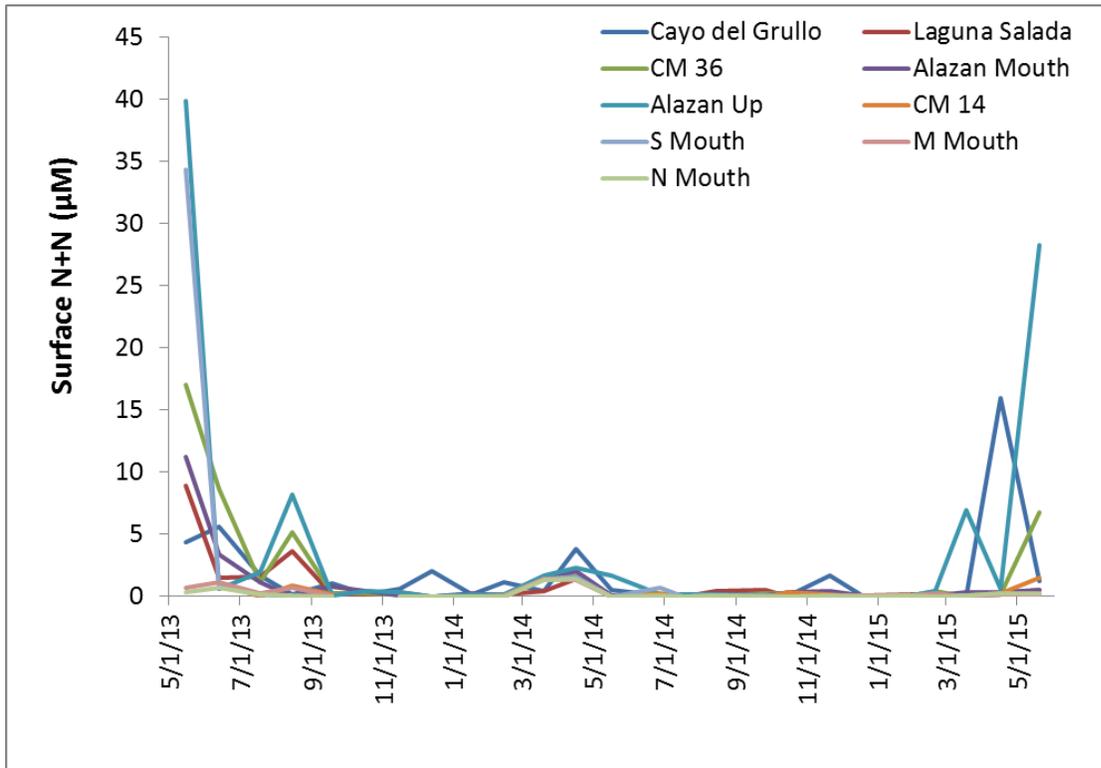


Figure 10. Surface N+N in Baffin Bay.

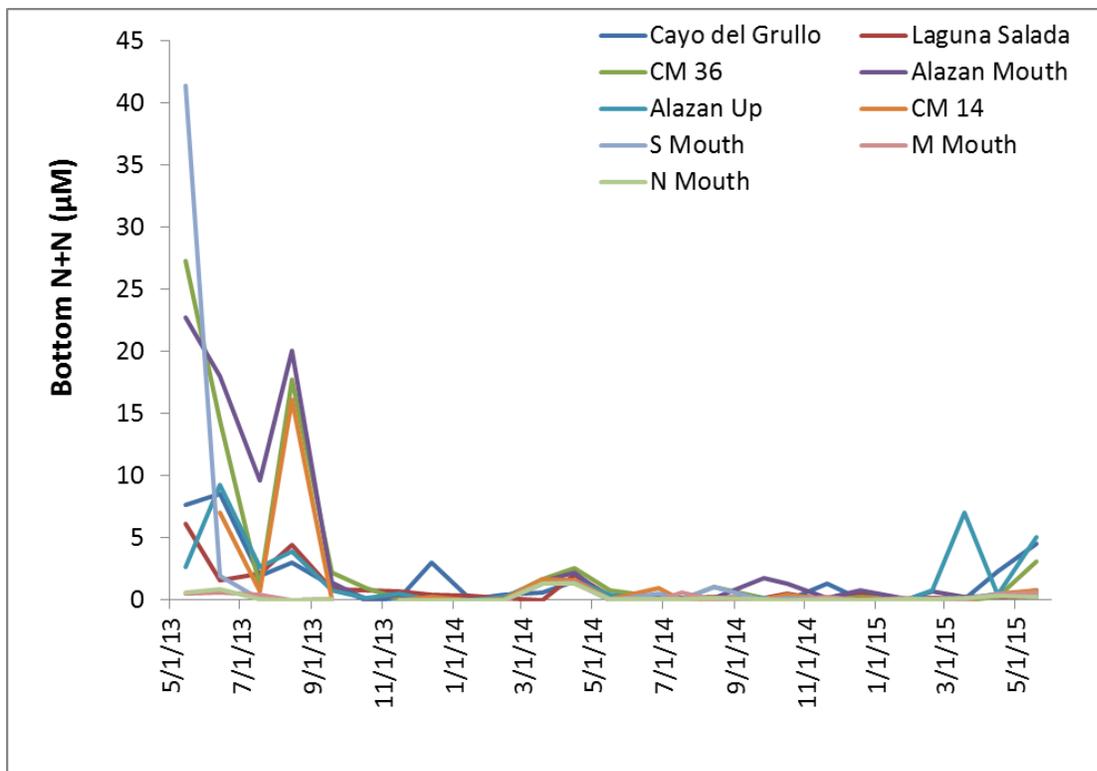


Figure 11. Bottom N+N in Baffin Bay.

observed at the upper Alazan and Laguna Salada sites, and declined thereafter through summer 2014 (Fig. 12). Moderately high ammonium concentrations (up to 7 μM) were observed again from June-August 2014, primarily in western Baffin Bay (Fig. 12). Thereafter, ammonium concentrations were relatively low until May 2015, when very high ammonium concentrations were observed at the upper Alazan (23.2 μM), CM 14 (11.5 μM) and south mouth (14.5 μM)

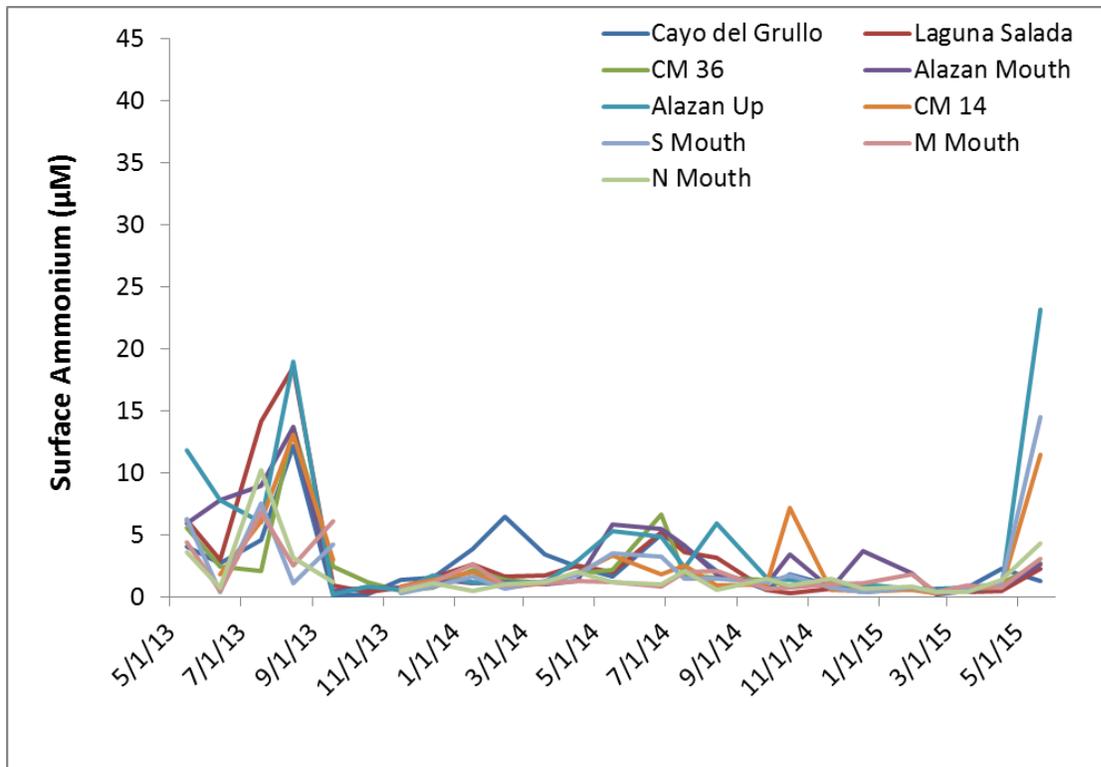


Figure 12. Surface ammonium in Baffin Bay.

sites. Elevated ammonium concentrations were also observed from May through August 2013 in near bottom water at most sites (Fig. 13). Ammonium concentrations in near bottom water declined thereafter and were generally low except for episodic increases. However, exceptionally high ammonium concentrations were once again observed in near bottom waters in May 2015, especially at the upper Alazan site (91.7 μM) (Fig. 13). By far, the dominant form of nitrogen during the study period was dissolved organic nitrogen (DON), with DON concentrations regularly exceeding 35 μM (Fig. 14). No clear seasonal pattern was observed in terms of DON. Surface orthophosphate concentrations exceeded 1 μM at the beginning of the study period at all sites except the north and central mouth sites ($< 0.5 \mu\text{M}$) (Fig. 15). Very high surface orthophosphate concentrations ($>5 \mu\text{M}$) were noted at CM 36, upper Alazan, and south

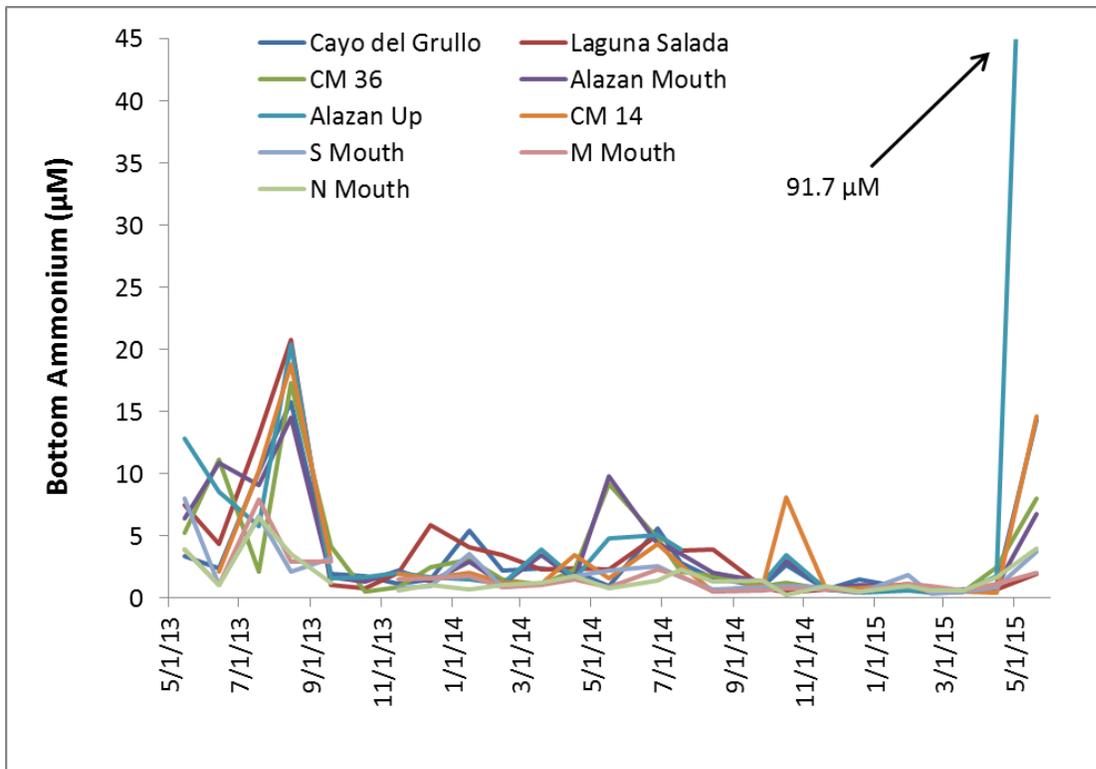


Figure 13. Bottom ammonium in Baffin Bay.

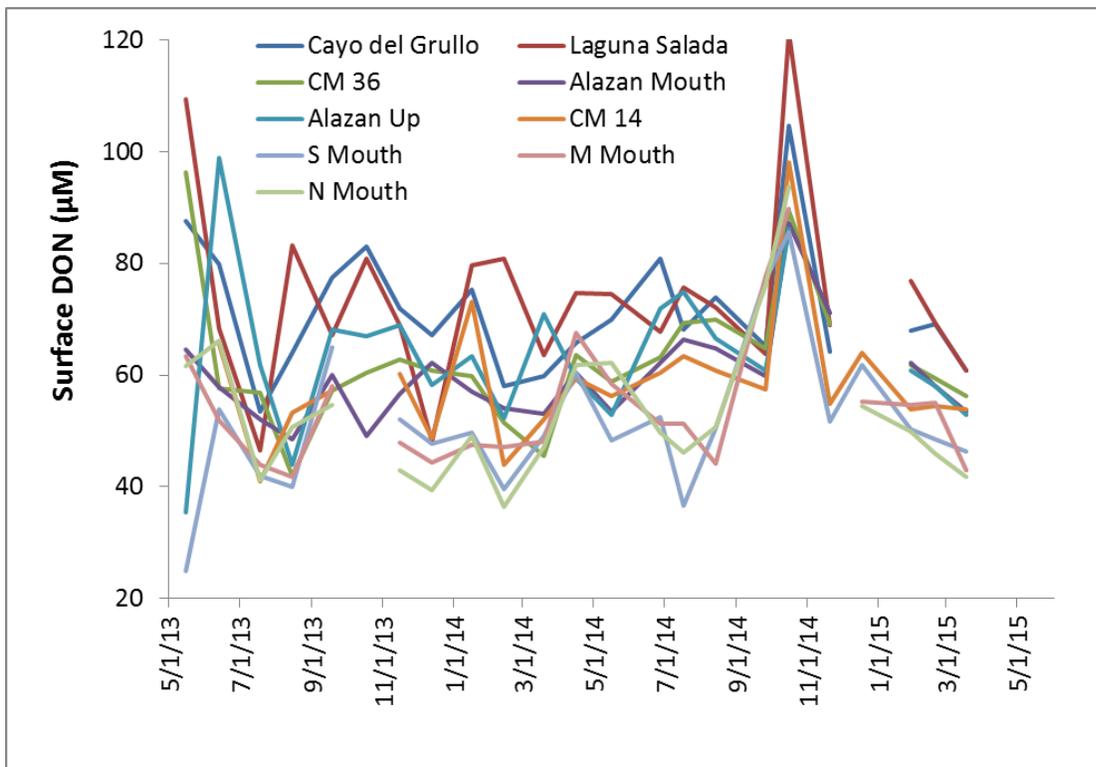


Figure 14. Surface dissolved organic nitrogen in Baffin Bay.

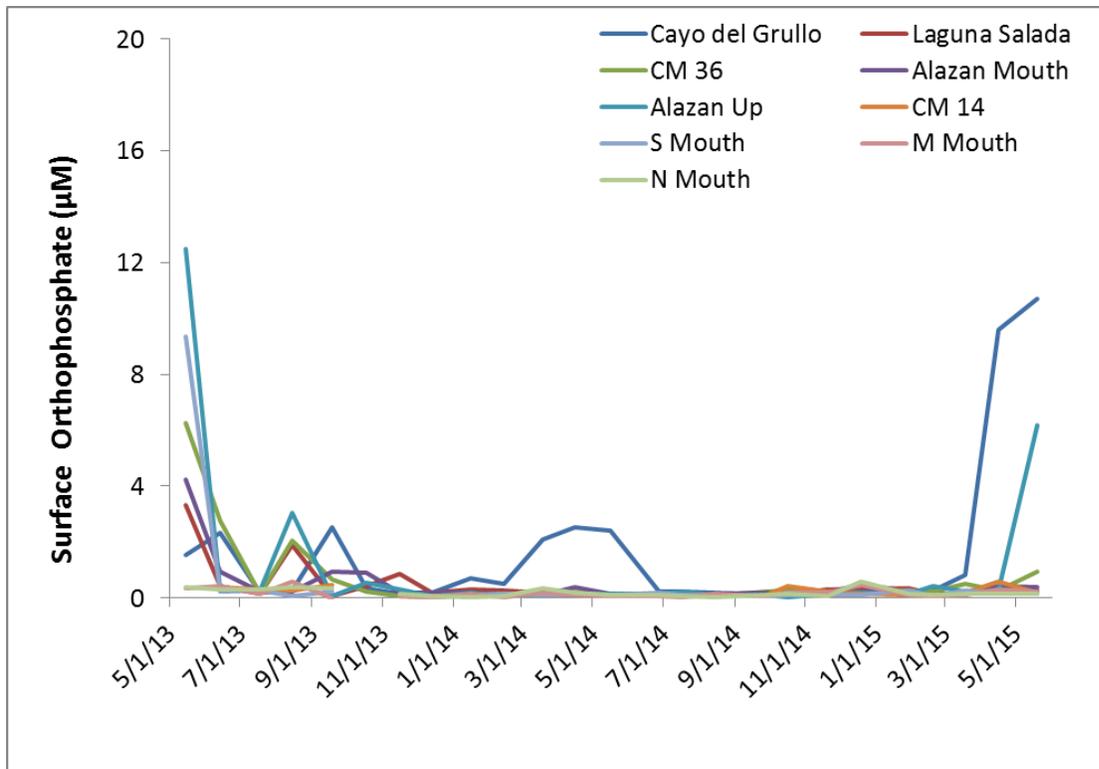


Figure 15. Surface orthophosphate in Baffin Bay.

site at the mouth (Fig. 15). Orthophosphate concentrations decreased after May 2013 and were generally $<1 \mu\text{M}$ thereafter through April 2015, with the exception of a small secondary peak in July-August 2013 in western Baffin Bay and concentrations of $2.1\text{-}2.6 \mu\text{M}$ in Cayo del Grullo from March-May 2014 (Fig. 15). In April-May 2015, very high concentrations were observed in Cayo del Grullo, ranging from $9.6\text{-}10.7 \mu\text{M}$ (Fig. 15). Likewise, in May 2015 the orthophosphate concentration at the upper Alazan site was $6.2 \mu\text{M}$ (Fig. 15). Orthophosphate concentrations were periodically elevated in near bottom waters, especially in May, June and August 2013 at various sites throughout Baffin Bay, and April-May 2015 in Cayo del Grullo (Fig. 16). The inorganic nitrogen to phosphorus ratio, one indicator of the limiting nutrient (i.e., N or P), displayed P-limiting conditions ($\text{N:P} > 16$) in summer 2013 and 2014 as well as May 2015, but displayed N-limiting conditions ($\text{N:P} < 16$) the remainder of the time (Fig. 17). Silicate concentrations were generally $>20 \mu\text{M}$, with the exception of sites at the mouth where concentrations as low as $<1 \mu\text{M}$ were occasionally observed (Fig. 18).

Dissolved organic carbon (DOC) represents a biogeochemically important constituent of coastal waters, where it supports bacterial respiration. DOC concentrations were exceptionally

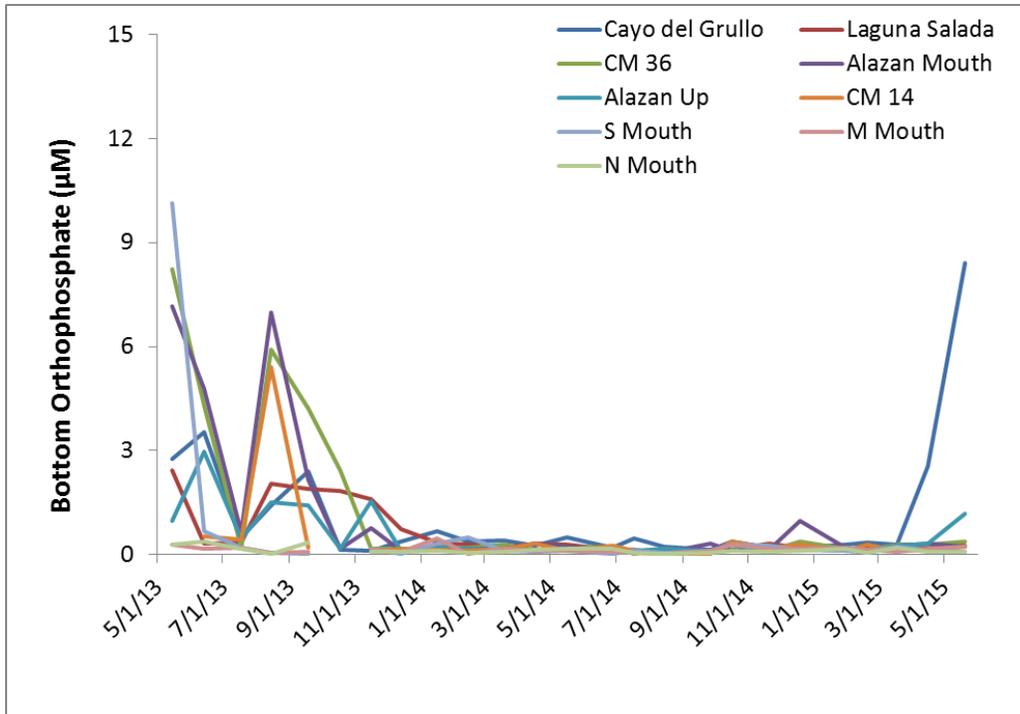


Figure 16. Bottom orthophosphate in Baffin Bay.

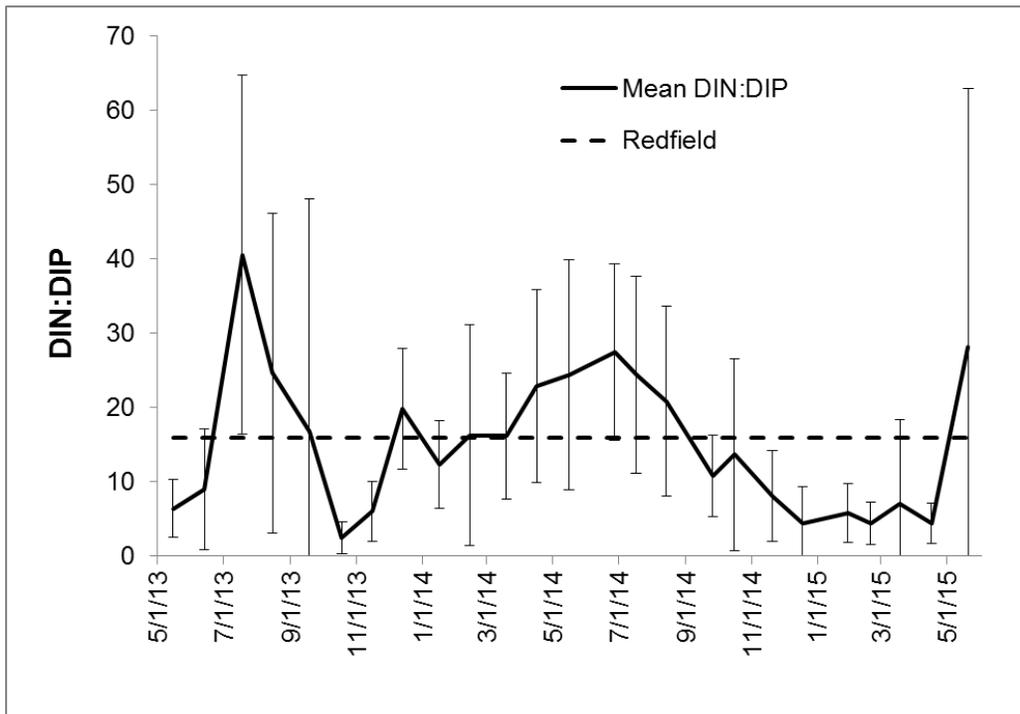


Figure 17. Ratio of dissolved inorganic nitrogen to orthophosphate in surface waters of Baffin Bay. Dashed line at N:P of 16 is the Redfield ratio, above which P is believed to be limiting to algal growth and below which N is believed to be limiting.

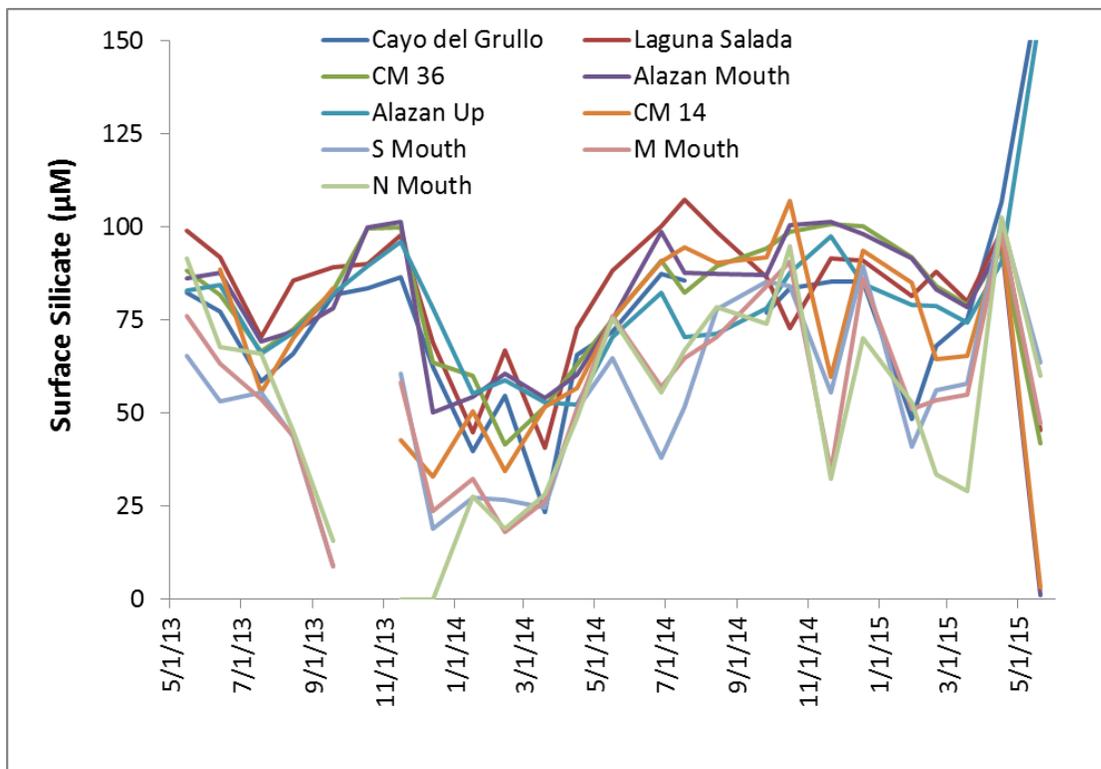


Figure 18. Surface silicate in Baffin Bay.

high at the beginning of this study, exceeding 1000 μM at all locations (Fig. 19). DOC subsequently decreased at most locations and remained fairly constant after summer 2013. Highest DOC concentrations tended to be found in the western portion of Baffin Bay.

Dissolved oxygen displayed a clear seasonal pattern that can be linked to temperature, with lowest levels being observed in the warmer months and highest levels in winter (Fig. 20). In summer 2013, several instances of sub-hypoxic bottom waters were observed. Yet in 2014, despite similarly high water temperatures, hypoxia was not observed. No clear spatial patterns are evident (Fig. 20).

Discussion

Results from the first two years of a multi-year water quality data collection effort in Baffin Bay show the presence of spatial-temporal patterns in terms of water quality dynamics in the system. Ultimately, this data, in conjunction with ongoing collections over the next two years and a reanalysis of historical TCEQ data should provide a comprehensive understanding of water quality conditions, as well as environmental drivers that affect water quality in Baffin Bay.

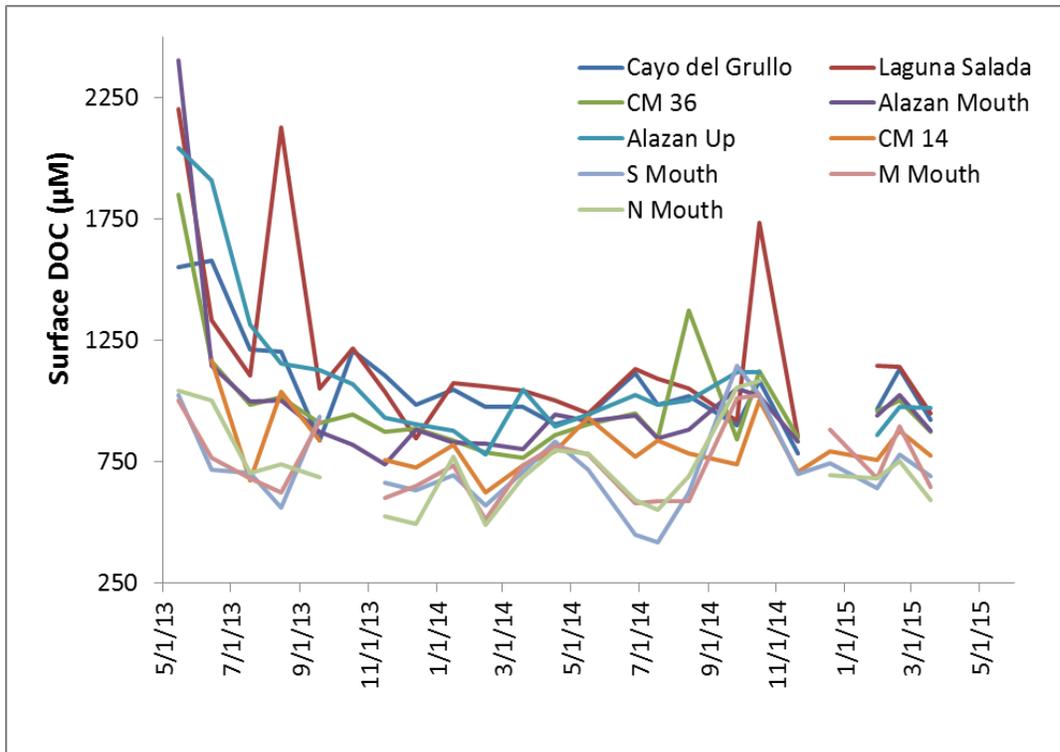


Figure 19. Surface dissolved organic carbon in Baffin Bay.

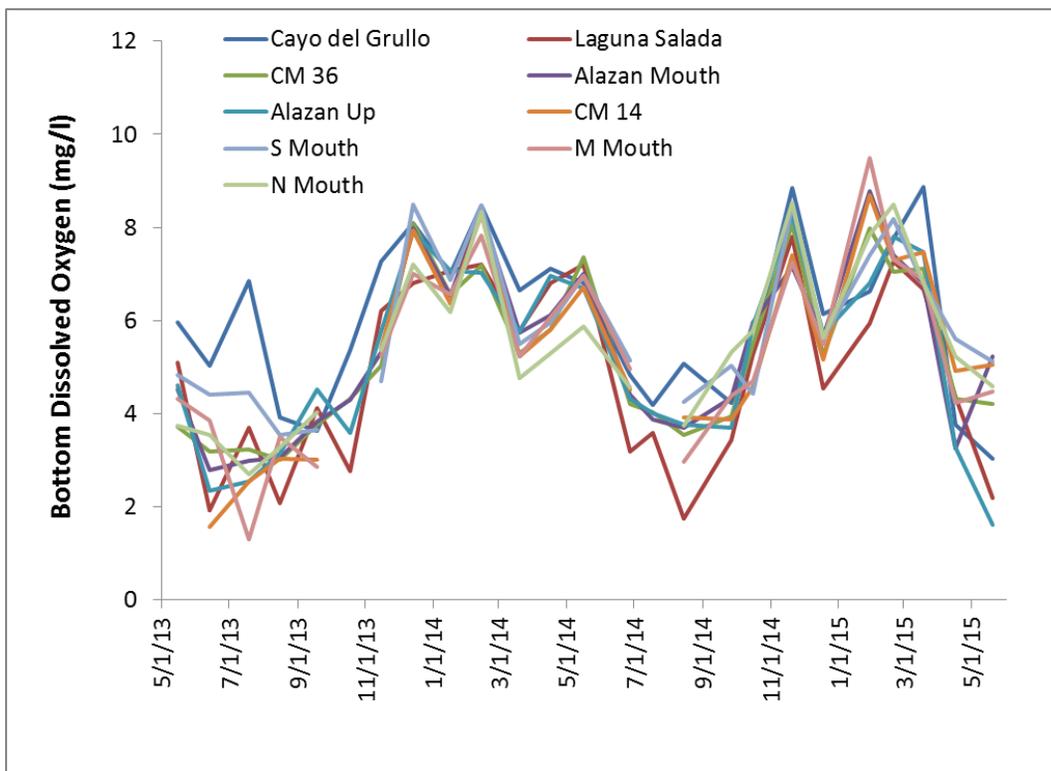


Figure 20. Bottom dissolved oxygen in Baffin Bay. Dashed line indicates the hypoxia threshold.

At the beginning of the study period, Baffin Bay was experiencing a significant, prolonged drought and concurrently a major bloom of the brown tide organism *Aureoumbra lagunensis* (Wetz, unpubl. data). Hypersaline conditions associated with drought have previously been shown to favor brown tide blooms in the system (e.g., Buskey et al. 1997, 2001). Chlorophyll *a*, a key indicator of phytoplankton (algal) biomass, exceeded TCEQ criteria throughout much of the study period and was frequently at levels that would be considered excessive even by National Coastal Condition Report (EPA, 2012) standards. The strong seasonal pattern of high chlorophyll in spring-summer can be explained in part by light and temperature. That is, phytoplankton growth is dependent on ample light levels for photosynthesis, and is often correlated with water temperature as well. Thus the winter low levels of chlorophyll may be indicative of either low light, low temperatures, or both. As a case in point, chlorophyll levels were nearly 2-fold higher in winter 2014-2015 compared to winter 2013-2014, which may be a result of water temperatures being 5°C warmer in 2014-2015.

From early fall 2013 through spring 2015, precipitation patterns developed that were more in accordance with long-term monthly averages, and salinities decreased. Despite the lower salinity levels, very high chlorophyll levels were noted in mid-2014 as in mid-2013. Only recently has chlorophyll noticeably decreased concurrent with several heavy precipitation events in spring 2015. Although speculative at this point, it is likely that the recent lower than expected chlorophyll levels may be attributed to significant flushing as well as reduced light from high turbidity that accompanied the spring rains in Baffin Bay. A longer-term goal of this study is to examine how phytoplankton (incl. brown tide) respond to the lower salinity, “wet” conditions that have developed and are expected to continue for the next months-year as a result of El Niño. An alternative explanation is that microzooplankton grazing may have been depressed during the hypersaline conditions, but have become important again with lower salinities (e.g., Buskey et al. 1997, 2001). Our group recently completed several experiments to quantify microzooplankton grazing during the recent lower salinity conditions and will be in a position to address this issue at a later date.

An interesting feature in terms of chlorophyll, observed in both 2013 and 2014, is a sharp decrease in chlorophyll between May and June in the eastern portion of Baffin Bay. In both years, this appeared to correlate with an intrusion of lower salinity, oligotrophic water from Laguna Madre in Baffin Bay. We believe this is due to the seasonal shift in wind direction along

the Texas coast, which results in southeasterly winds becoming established during this time of year. Interestingly, salinity and chlorophyll levels were largely unaffected in the western Baffin Bay. A TAMU-CC Ph.D. student (Emily Cira) will be conducting a study of this phenomenon beginning in 2015. If confirmed, this finding highlights the importance of water exchange with Laguna Madre in terms of water quality dynamics in Baffin Bay.

In addition to light and temperature, nutrient (primarily nitrogen and phosphorus) availability is a major control on phytoplankton growth. Results from this study show that the nitrogenous nutrients available to phytoplankton are primarily DON and secondarily ammonium. During this study period, surface N+N levels were relatively low except in May 2013 and at select locations in March-May 2015. As in other estuarine systems, it appears that nutrient cycling and availability is influenced by water temperature in Baffin Bay. For example, during summer 2013 and 2014, relatively high concentrations of ammonium were prevalent compared to oxidized N+N, and may result from such processes as fluxes out of near bottom waters, recycling within the water column, and photochemical alterations, among others. Furthermore, elevated concentrations of ammonium were observed during summer in near bottom waters, consistent with studies from this and other systems showing release of reduced nitrogen from suboxic sediments under warm conditions (e.g., An and Gardner 2002). These reducing conditions became less prevalent as temperatures decreased during fall and winter, and consequently ammonium levels in both surface and bottom waters decreased considerably. Despite this, DON concentrations remained elevated throughout the year. In fact, the total dissolved nitrogen (i.e., DON + ammonium, N+N) and DON concentrations observed in Baffin Bay are consistently higher than many other estuaries in the Gulf of Mexico, including those of the central Texas coast (e.g., Bianchi 2007; Mooney & McClelland 2012; Wetz unpubl. data). Prevalence of high concentrations of reduced nitrogen such as ammonium and DON are important because they have been implicated as potentially favoring dominance by the brown tide organism over other healthy phytoplankton (Gobler et al. 2013). Orthophosphate concentrations in surface waters were only episodically elevated, though like ammonium, high near bottom water concentrations were occasionally observed during summer. Our group has collected additional samples to determine total dissolved phosphorus and dissolved organic phosphorus concentrations, and will report that data in future syntheses. A previous study (Cotner et al. 2004) highlighted the potential for phosphorus to become limiting (to phytoplankton growth) in Baffin Bay, and our

initial findings based on inorganic nitrogen to phosphorus ratios supports this notion. As part of a GLO-funded study, we began conducting nutrient addition bioassays in late winter 2015 to further test for nutrient limitation of phytoplankton growth in Baffin Bay.

A longer term goal of this study is to begin to identify sources of nutrients to Baffin Bay. Clearly the sediments in this system are a potentially important source of nutrients to the water column, at least on a seasonal basis. At this point, the role of watershed derived nutrients remains enigmatic and will require additional data collection. However, the sharp increase in nutrient concentrations in both the upper Cayo del Grullo and Alazan Bay after recent heavy rain events points to watershed source(s). Another study has shown that nitrate inputs to Baffin Bay may be rapidly converted to ammonium as a result of the DNRA process (An and Gardner 2002). Thus it is possible that the effects of smaller rainfall events on nutrient inputs/concentrations may be masked by this and other similar processes. Other sources of nutrients, such as groundwater and atmospheric deposition, must be considered in future nutrient budgeting efforts as well.

Organic matter concentrations in Baffin Bay tended to be very high during the study period. Sources of DOC are unclear, though DOC concentrations tended to be much higher in the western portion of Baffin Bay, possibly indicating tributary sources and/or internal sources such as phytoplankton exudation. Additionally, DOC concentrations were occasionally high in near bottom waters (data not presented). High particulate organic matter concentrations were also observed and tended to mirror chlorophyll levels, indicating significant contribution of phytoplankton carbon to water column particulate organic matter. These high levels of organic matter are important because they may fuel microbial respiration and biological oxygen demand. Near bottom oxygen levels showed a distinct seasonal cycle that is undoubtedly temperature related, with lower temperatures capable of holding more oxygen than higher temperatures. Nonetheless, we routinely observed undersaturated oxygen levels, especially during the summer, indicative of intensive microbial respiration and utilization of labile organic matter. Dissolved oxygen occasionally reached hypoxic (<2 mg/l) levels and were generally <4 mg/l during summer. Previous studies have shown that hypoxic dissolved oxygen levels, and in some cases oxygen levels of <3-5 mg/l, can have sublethal and/or lethal effects on benthic organisms (e.g., Ritter and Montagna 1999; Diaz and Rosenberg 2008). Nonetheless, monthly sampling is clearly insufficient to quantify dissolved oxygen “mean state” as well as timescales of change in

a dynamic system such as Baffin Bay. Consequently, our lab has been deploying sondes along the longitudinal axis of Baffin Bay since February 2015 to characterize these aspects of dissolved oxygen dynamics in Baffin Bay.

In the third year of this study, we will continue to focus on additional data collection, which will be crucial given the recent sharp drop in salinity and increase in nutrient concentrations in Baffin Bay. This additional data will be the first of its kind to quantify the influence of wet versus dry periods on water quality in Baffin Bay. Additionally, we will continue to deploy Hydrolab sensors in Baffin Bay for continuous monitoring of water quality, allowing for better determination of daily to weekly timescales of water quality change. Finally, research will be undertaken to explore in much greater detail the interactions between physical drivers (circulation, winds) and water quality dynamics in the system.

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